

# Temporal and Spatial Variation of Nutrient Leaching from Agricultural Land in Latvia: Long Term Trends in Retention and Nutrient Loss in a Drainage and Small Catchment Scale

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**Abstract** - Agricultural run-off monitoring and losses of nitrogen (N), phosphorus (P) from small (60 – 960 ha) catchments and drainage fields (12 – 77 ha) in Latvia were analyzed. The catchments represent areas with different agricultural production intensity, landscape, soils and climate. Results from the long-term monitoring show (i) large differences in levels of losses between the catchments and fields with different farming intensity, (ii) large annual and interannual variability due to climate impact and soil conditions, (iii) few time trends in nutrient losses. These results are discussed in relation to pollution from non-point and point sources. Besides, these data demonstrate changes in agriculture driven by large-scale changes in land-use intensity. However, despite the increased implementation of mitigation methods by farmers, improvements in water quality at the catchment scale can take a long time. Long-term monitoring programmes to detect trends are required.

**Keywords** – Agriculture, monitoring, nutrients

## I. INTRODUCTION

High nutrient levels in the surface waters can result from several agricultural pollution sources [1, 6, 7, 8, 10, 11, 14, and 26]. Non-point (diffuse) agricultural pollution or agricultural run-off is a major issue causing the pollution of surface water ecosystems in Latvia [12, 13, and 26]. The pollution of surface water due to nutrient input from diffuse sources is not easy to observe, measure and mitigate [23, 25, and 26]. Non-point agricultural pollution occurs everywhere and always, and agro-environmental measures to reduce emissions in agriculture can provide limited results. A major cause of point source agricultural pollution comes from manure management arising from large-scale animal farming. Point source pollution is commonly associated with surface run-off from farm territories, manure storage, leakage from manure and urine reservoirs and other sources of poor manure management practices [1, 7, 9, and 24]. Therefore, in an assessment of agricultural run-off, it is crucial to be able to estimate the impact of animal farming and diffuse pollution on water quality by an apportionment of both agricultural pollution sources from the other types of pollution. In order to comply with EU regulations e.g., Nitrate Directive (ND) [22] and Water Framework Directive [3], it is important to understand the effect of the leaching of nitrates and other plant

nutrients on eutrophication phenomenon. Monitoring of water pollution by agricultural run-off in Latvia has been implemented by the Department of Water Management and Environmental Engineering of Latvia University of Agriculture since 1994 [12, 15].

## II. MATERIAL AND METHODS

### A. Monitoring of diffuse agricultural pollution

An assessment of diffuse agricultural pollution has been implemented in three monitoring stations and four sampling points (Fig. 1). Diffuse pollution monitoring stations are placed in three small catchments (river basins < 10 km<sup>2</sup>). The Bērze, Mellupīte and Vienzienmīte sites are located in different parts of Latvia and represent regions with different climate, soil, slopes, crops and farming intensity [8, 10, 14]. There are no other non-agricultural pollution sources located in these catchments. The soils at the monitoring sites are imperfectly to poorly drained. Most of the agricultural land in the small catchments is drained with tile drains (depth 1.1 – 1.3 m, spacing between drains 10 – 32 m).

However, monitoring of agricultural run-off in regions with humid climate is performed at several geographical scales. Nutrient losses from arable land might be measured at a drainage field level. Field scale run-off represents a combined effect on the water quality from farming practice, crop rotation, application of fertilizers, etc.



Fig. 1. Agricultural pollution monitoring stations and monitoring points in Latvia.

The integrated influence on nutrient run-off of variations in farming practices, erosion, soil and topography within the small drainage basin (catchment) might be studied in a better way than on the field scale. In a small catchment scale, emission rates (loads) that can contribute to the nutrient enrichment of the surface water ecosystem can be examined. The method complies with the EU ND Monitoring Guidelines [5, 19, 20], the internationally accepted practice [26, 27] and methods proposed in Latvia [19]. A description of the monitoring stations and monitoring points in Latvia can be found in Table 1. A monitoring station is a monitoring site equipped with a hydraulic structure to discharge measurements with data loggers. Permanent stations with structures and advanced equipment are used for diffuse pollution assessment in the Lielupe (Bērzes station), the Venta (Mellupīte station) and the Gauja (Vienziemītes station) river basin districts since 1994. For the apportionment of the point and non-point (diffuse) pollution sources, Skrīveri (since year 2000), Bauska (since year 1995) and Auce (since year 2004) monitoring points have been established on the small catchment scale. Monthly grab samples are collected at the monitoring points, however the discharge is modeled.

Specific hydraulic structures, i.e. Crump, V-shape Crump and combined profile weirs and data loggers for continuous and automatic measurements have been used in Bērze, Vienziemīte and Mellupīte monitoring stations. Data loggers measure the water level on the weir and calculate discharge continuously. Generally, automatic flow proportional sampling triggered by data logger was used in the stations. Based on a flow proportional sampling procedure, monthly composite water samples were collected. The grab sampling method used in the monitoring points have lower statistical power because the sampling frequency and method is less sufficient to cope with the high variability of water quality [5, 18] and the quick response to local meteorological conditions.

The parameters analysed included total N,  $\text{NO}_3\text{-N}$ ,  $\text{NH}_4\text{-N}$ , total P,  $\text{PO}_4\text{-P}$  and pH. Water analyses were performed according to the standard methods.

The landscape of the Bērze site is flat lowland and 98 % of the catchment soils are cultivated. Due to natural high soil fertility, winter wheat has become the main crop in the Bērze catchment. The share of arable crops increased up to 80-90 % during 1994 - 2010. Farmers are using modern equipment, and rather intensive technology for Latvia conditions. The Bērze catchment is characterised by relatively intensive crop production and modern technologies, as compared to the present average conditions in Latvia, e.g. a fertilizer application in few fields reached  $170 \text{ kg N ha}^{-1} \text{ year}^{-1}$  in 2010. However, by contrast, farms in the Vienziemīte catchment apply low amounts of fertilizers (on average  $4 \text{ kg N ha}^{-1} \text{ year}^{-1}$ ). The Vienziemīte site is a typical example of low input farming and could be used as a reference site for the assessment of water quality in agricultural catchments. The Mellupīte catchment represents average farming conditions and could be considered typical for present agriculture in Latvia. Three large farms are using intensive agricultural technology, whereas a few farms produce only for self-

consumption with low fertilization rates and without pesticides.

TABLE I  
DESCRIPTION OF THE AGRICULTURAL POLLUTION MONITORING SITES (DIFFUSE POLLUTION)

Site, monitoring, level for water sampling	Area, ha	% cultivated	Main soil type *	Description of agriculture
<b>Monitoring stations</b>				
Vienziemīte, Cēsu district				
Small catchment	592	78	Chromic Cambisol Sandy loam	Low input farming
Drainage field	67	100		Low input farming
Mellupīte, Saldus district				
Small catchment	960	69	Stagnic Luvisol Loam, clay loam	Moderately intensive farming
Drainage field	12	100		Intensive grain farming
Bērze, Dobeles district			Calcric Cambisol Inceptisol Silty clay loam	Intensive grain farming
Small catchment	368	98		Intensive grain farming
Drainage field	77	100		Intensive grain farming
<b>Monitoring points</b>				
Vecauce, Dobeles district			Gleyic Cambisol Sandy loam	Intensive grain farming
Small catchment	60	90		
Bauska, Bauskas district.			Chromic Cambisol Silt loam	Intensive farming, mixed crops
Small catchment	750	95		
Skrīveri, Aizkraukles district			Uutric Podzoluvisol Silty clay loam	Moderately intensive farming
Small catchment	890	40		

\* FAO soil classification

#### B. Monitoring of the point source agricultural pollution

Monitoring of the point source agricultural pollution in Latvia was implemented in three small catchments, where pollution from large livestock farms has an impact on water quality [1, 7, 9, and 24]. The Bauska, Vecauce and Ogre sites are characterized by large applications of animal manure (slurry). The catchments were previously a part of large state and collective farms that specialized in pig production. A description of the point source pollution monitoring sites is presented in Table 2.

The Vecauce (LLU Vecauce, "Pūpolu" farm) farm started pig production in 1990 and produces about 1 000 – 2 000 pigs per year today. In the upper part of the Vecauce farm catchment (60 ha), forest and wetlands cover 5 ha (9 %). 30 ha of the catchment area are used for slurry application. On average, 200 m<sup>3</sup> slurry per ha and year have been applied on

winter grains during the vegetation seasons from 1995 - 2010. This application rate corresponds to a nutrient supply of approximately 180 - 360 kg N ha<sup>-1</sup> and 13-26 kg P ha<sup>-1</sup>. Previous experience demonstrated that the main principles of good agricultural practice are used in manure handling in the Vecauce farm.

The Bauska ("Strautu") farm was established in 1970 and reached full production (12 000 fattening pigs per year, 55 000 m<sup>3</sup> pig slurry per year) in 1976. Until 1987 tractor-driven tankers were used for slurry application. In 1987 a slurry irrigation system (226 ha) was installed. Today the pig production is 8 000 – 10 000 fattening pigs per year, but the slurry utilization area is only 50 hectares, due to changes in land use after privatization. Slurry has been applied on grassland fields throughout the whole year. On average 900 m<sup>3</sup> of slurry has been applied per ha annually, representing a nutrient supply of approximately 630 kg total N ha<sup>-1</sup> and 80 kg P ha<sup>-1</sup> that does not meet the requirements of ND (170 kg

N ha<sup>-1</sup>). Simultaneous grab sampling is performed downstream from the farm and in the upstream part of the catchment to assess the impact of point agricultural pollution.

The Ogre farm was completely closed after 15 years of production in 1992. The farm produced about 30 000 pigs per year prior to its closure. Earth lagoons were used for slurry storage and still remain full of slurry at this farm. Soil in the surrounding territory of the Ogre farm is still polluted due to improper manure handling during 1977 - 1991. Slurry was applied by tractor tankers and sprinkler irrigation on 240 ha of land and several leakage accidents took place from the lagoons. Old pollution still has an impact on water quality in that catchment [1].

The measurements in Bauska, Ogre and Vecauce were based on manual water sampling at regular intervals (once per month) since year 1995.

TABLE II  
DESCRIPTION OF THE AGRICULTURAL POLLUTION MONITORING SITES (POINT SOURCE POLLUTION)

Site, monitoring level	Area, ha	Cultivated land, %	Soil	Description of agriculture
Auce, monitoring point				
Small catchment in slurry application area	60	90	Sandy loam	Intensive grain farming, arable land 80 %. Slurry application on 30 ha within catchment.
Bauska monitoring point				
Small catchment in slurry application area including pig farm	800	95	Silt loam	Intensive farming, mixed crops. Slurry dumping site on 50 ha and pig farm within catchment.
Ogre monitoring point				
Small catchment in slurry application area including pig farm	300	25	Silty clay loam, Silt loam	Moderately intensive farming, old slurry lagoons and polluted territory of former pig farm within catchment, farm closed in 1992.

### C. Climate and Hydrology

Latvia is situated in a humid and moderately mild climatic region where rainfall exceeds evaporation, resulting in percolation losses from the soil during spring and autumn. The annual precipitation in Latvia is 550 - 750 mm per year. A wide range of precipitation from 1994 - 2010 was characterized by dry years in 1996, 2006, normal years in 1994, 2004, and wet years in 2007 and 2010.

Similarly, maximum and minimum extremes were observed in the water discharge (Fig. 2) in the monitoring catchments during the period of 1994 - 2010. The year 2006 was the driest since 1974 (run-off 116 mm per year in the Vienziemīte site) and 1995 was a wet year with the run-off 467 mm per year at the same Vienziemīte monitoring site. The wide range of climatic and hydro meteorological

conditions at the Bērze and Mellupīte sites was illustrated by a run-off ranging from less than 100 mm to more than 300 mm [2, 3, and 26].

To obtain values for the runoff from the manure utilization points and the Mellupīte catchment area, the METQ96 model was used. The model contains sub routines for snow accumulation and ablation, soil moisture accounting, groundwater and capillary fringe accounting, runoff generation and a simple routing procedure. Input parameters to the model are daily values for temperature, precipitation and the vapor pressure deficits which are standard meteorological data that are available in meteorological stations in Latvia. A detailed description of the model is given by A. Zīverts et al. [16].

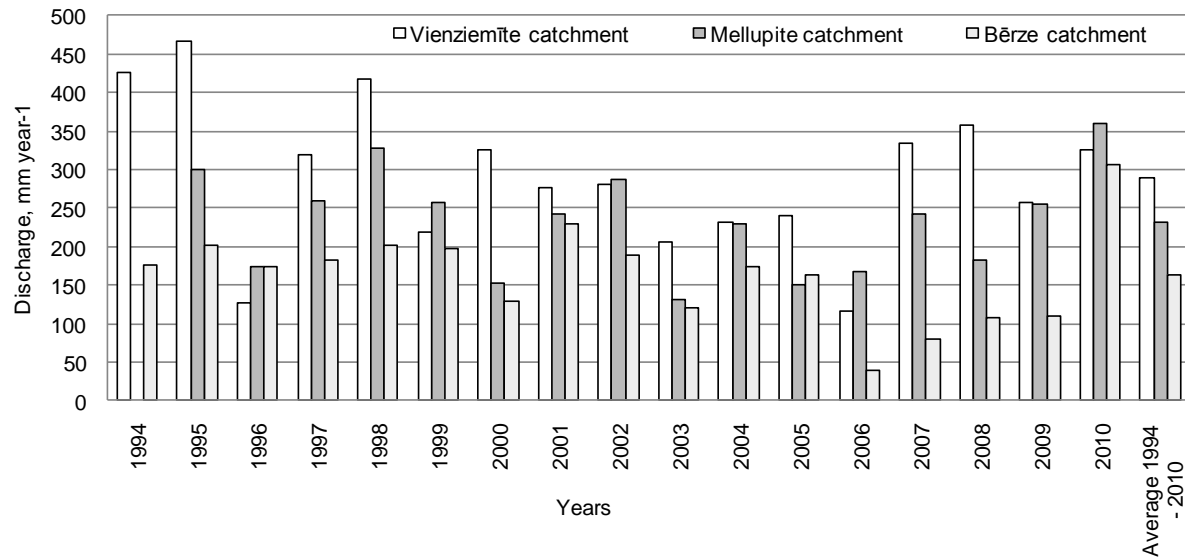


Fig. 2. Water discharge in small catchment scale in the monitoring station sites.

#### D. Nutrient run-off

Nutrient leakage was calculated by multiplying the nutrient ( $N_{tot}$ ,  $P_{tot}$ ) concentrations of the composite water samples with the total volume of water discharged during the corresponding water sampling period. The research in the small catchment scale demonstrates that the rapid response of discharge due to rainfall or snowmelt and fast changes in water quality parameters caused nutrient enrichment on a local scale in small catchment or drainage field are observed more often than at the large river scale. Usually, during the flood period more frequent water sampling could be recommended, as mentioned in ND [17, 19, 22].

In order to calculate of nutrient run-off for the corresponding measurement period, the daily nutrient concentrations ( $C_i$ ) are interpolated from monthly nutrient values [19].

$$V = 0,01 \sum C_i * Q_i \quad (1)$$

where

$V$  - nutrient run-off,  $kg\ ha^{-1}\ year^{-1}$ ;

$C_i$  - average daily concentrations,  $mg\ l^{-1}$ ;

$Q_i$  - average water discharge,  $mm$ .

Hourly and daily discharges ( $Q_i$ ) are calculated automatically by a data logger, where water level measurements are performed every 2 - 3 minutes. Moreover, the data logger calculates the volume of the water discharged over the hydraulic structure and after bypass of certain amount of water sends a signal to the water sampling pump. Our experience shows that the logger-triggered sampling frequency could be 10 - 15 subsamples per day. One water proportional composite sample consists of a large number of subsamples gathered during a period of one month, i.e., 12

water samples were collected, if there was discharge all year around.

### III. RESULTS AND DISCUSSION

#### A. Nutrient concentrations

EU ND monitoring guidelines [5] for the implementation of Article 3.1 propose that long term (20 - 30 years) nitrate trends have been analyzed for:

- Yearly average values;
- Winter period values;
- Maximal concentrations of nitrates.

The guidelines [5] suggest that the presence of nitrate nitrogen ( $NO_3^- - N$ ) during the winter period (October – March) better describes the potential amount of available nitrogen resulting in eutrophication in summer. When analyzing nutrient impact during the vegetation season, it is more useful to measure the total nitrogen ( $N_{tot}$ ).

However, EU Water directives and guidelines [4, 20, 21, 22] do not explain what values of nitrate concentrations should be used (e.g., average annual, winter or maximum peak nitrate) to compare the limit value for nitrates established by ND –  $50\ mg\ l^{-1}\ NO_3^-$  or  $11.3\ mg\ l^{-1}\ NO_3^- - N$  [5, 22]. Therefore, the assessment of nitrate concentrations was performed for maximum peak  $NO_3^- - N$ , winter concentrations of the  $NO_3^- - N$  and summer concentrations of the  $N_{tot}$ .

#### B. Nitrate concentrations

Limit values of nitrate ( $11.3\ mg\ l^{-1}$ ) established by the Directive were often exceeded in the small catchment and drainage run-off from fields with intensive farming in the Bērze monitoring site since the year 1996 (Fig. 3 and 4). Generally, drainage field run-off has higher nitrate concentrations and a higher risk that threshold values of ND could be exceeded.

Surprisingly, high nitrate values in the Bērze drainage field run-off were found in November 2006, when drainage run-off appeared after the dry summer - autumn period. As shown in Fig. 3, the nitrate concentration may be extremely high. The VSIA "Centre for Agrochemical Research" has analyzed mineral nitrogen in several fields. Data of the soil mineral nitrogen, presented in, provide the information on soil leakage potential in the autumn of 2006 resulting in high concentrations when drainage run-off started in November. Run-off transported a considerable amount of nitrogen that

was accumulated in the soil during the summer and early autumn due to the dry and hot weather conditions [13]. In addition, the soil had many cracks and macro pores. In the small catchment scale (Fig. 4), long-term concentration trends of different nitrogen compounds show an increase by 50 - 90% during 1994 - 2010.

Due to the extremely high nitrate values in 2006/2007 in the drainage field scale (Fig. 3), trends are impossible to evaluate. However, summer  $N_{tot}$  concentration values increased by 20%.

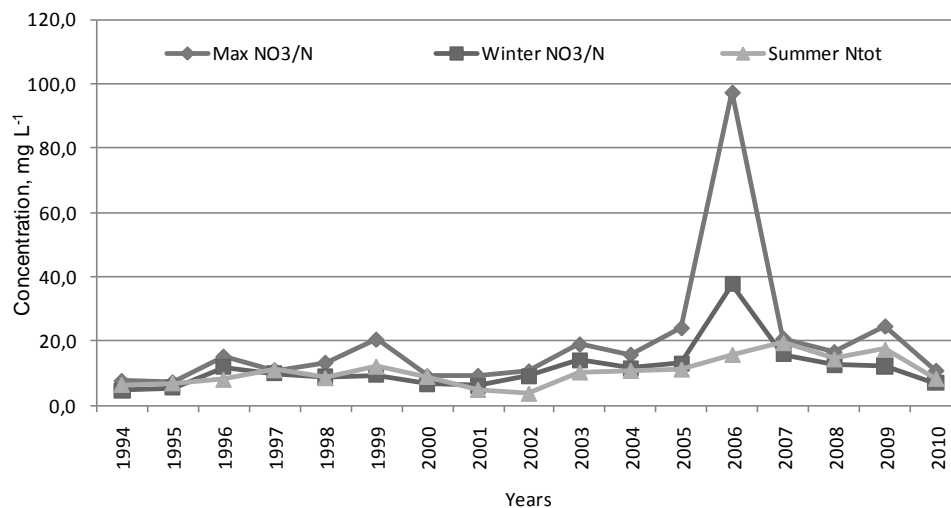


Fig. 3. Annual average nitrogen concentrations in drainage field run-off, Bērze monitoring station.

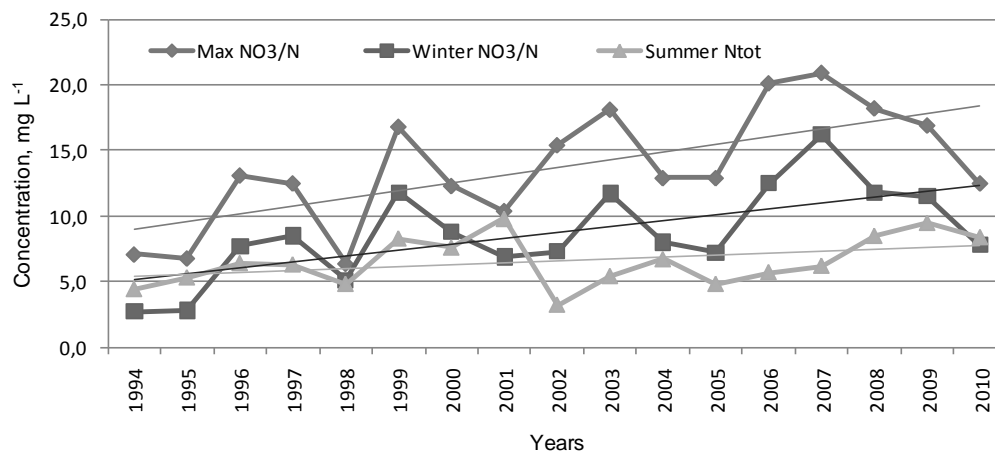


Fig. 4. Annual average nitrogen concentrations in small catchment run-off, Bērze monitoring station.

With regard to the Nitrate Directive, it should be noted that the limit values of nitrate ( $11.3 \text{ mg l}^{-1}$ ) established by the Directive were exceeded in the drainage field of the Mellupīte site during 2006 - 2008 (Fig. 5). Similarly to the Bērze site in that the field farmer used intensive technologies and farming methods. In the small catchment scale (Fig. 6) in the Mellupīte site, the limit concentrations of nitrates were exceeded only during 2003. Both on the catchment and the field scale, long-term nitrate concentrations show an upward trend (Fig. 6). In

the Vienziemīte site, which could be used as a reference site with low input agriculture, both on the catchment and the drainage field scale, during the period of 1994 - 2010, nitrogen concentrations close to the limit values of ND were not determined. Annual average concentrations in plot and field scale have a relatively small difference (Fig. 7 and 8), which could be attributed to normal variation. Long-term trends show no significant increase or decrease of nitrogen compounds in the Vienziemite site during 1994 - 2010.

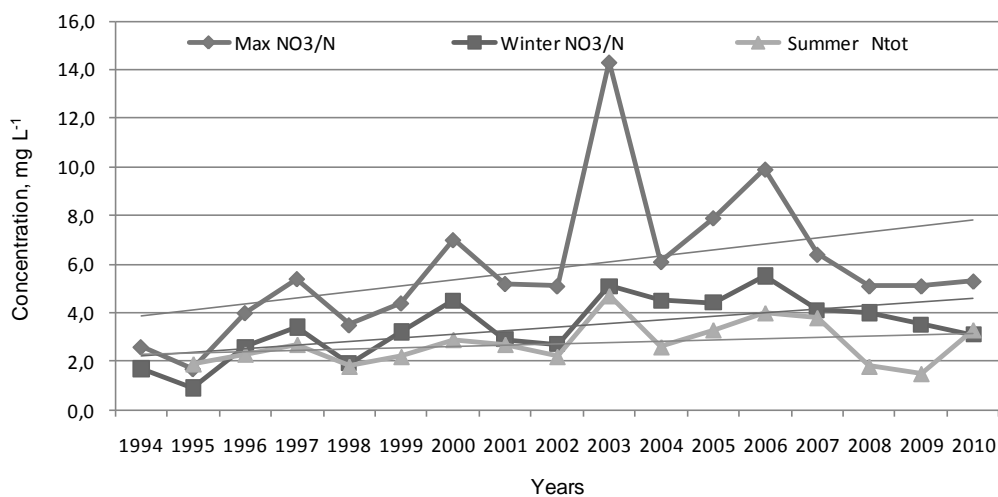


Fig. 5. Annual average nitrogen concentrations, Mellupīte drainage field.

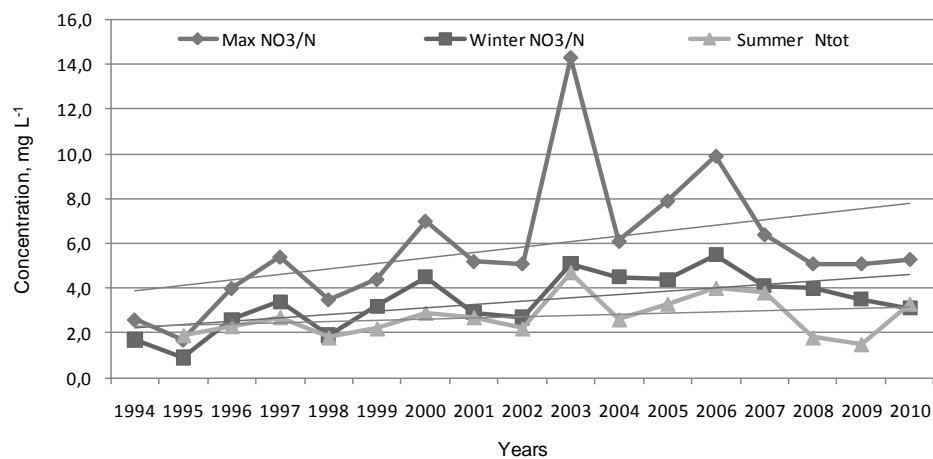


Fig. 6. Annual average nitrogen concentrations, Mellupīte small catchment.

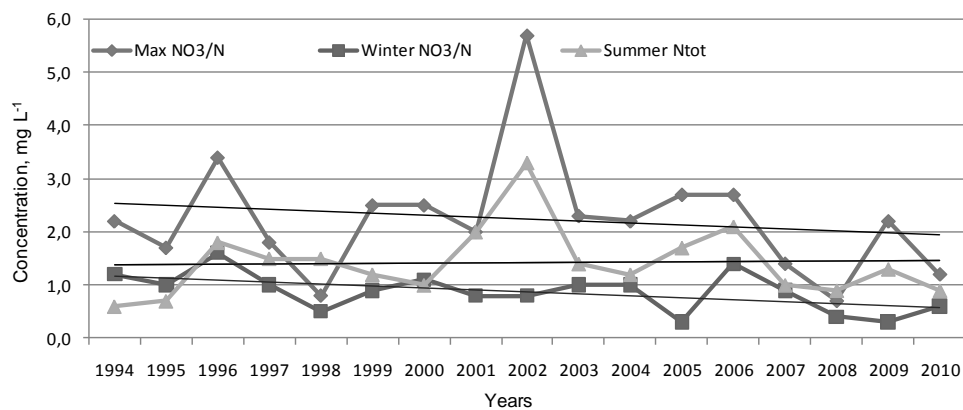


Fig. 7. Annual average nitrogen concentrations, Vienziemīte field drainage.

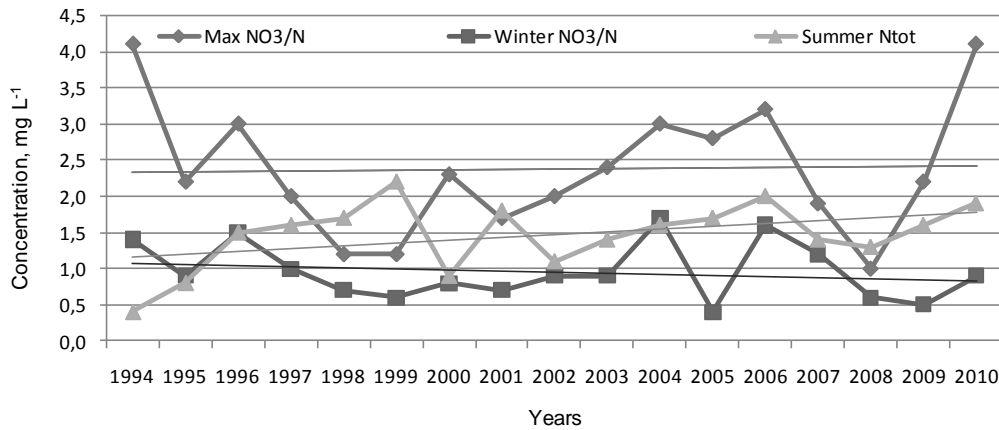


Fig. 8. Annual average nitrogen concentrations, Vienziemīte small catchment.

### C. Nutrient run-off in the diffuse pollution monitoring sites

Nutrient run-off from diffuse pollution monitoring sites is presented in Fig. 9, 10 and 11. Monitoring results have demonstrated the evident influence of water discharge on nutrient run-off. High precipitation rates determined continuous discharge on the small catchment scale during the wet years of 2007 and 2010. Even field drainage has a short period without discharge in June at the Bērze site. The results indicated that the highest nitrogen leakage was measured in the years 2007, 2008 and 2010. The research has shown that, despite high nitrogen concentrations in field drainage in the Bērze station, high nitrogen run-off was observed in the Mellupīte site. This was accompanied by an increase in discharge in the Mellupīte drainage systems. On the small catchment scale, relatively intensive farming in the Zemgale region (Bērze) contributed to a higher nutrient run-off (Table 3). Relatively low nutrient levels were measured in Vienziemīte, in the site with low input agriculture.

TABLE III

ANNUAL AVERAGE DISCHARGE AND NUTRIENT RUN-OFF FROM DIFFUSE POLLUTION SOURCES

Monitoring station	Discharge, mm	Run-off N <sub>tot</sub> , kg/ha year	Run-off P <sub>tot</sub> , kg/ha year
Vienziemīte monitoring station			
Small catchment, 592 ha	290	5.6	0.11
Field drainage, 67 ha	251	4.6	0.13
Mellupīte monitoring station			
Small catchment, 960 ha	233	10.2	0.12
Field drainage, 12 ha	246	17.9	0.13
Bērze monitoring station			
Small catchment, 368 ha	164	14.6	0.34
Field drainage, 76 ha	332	17.8	0.50

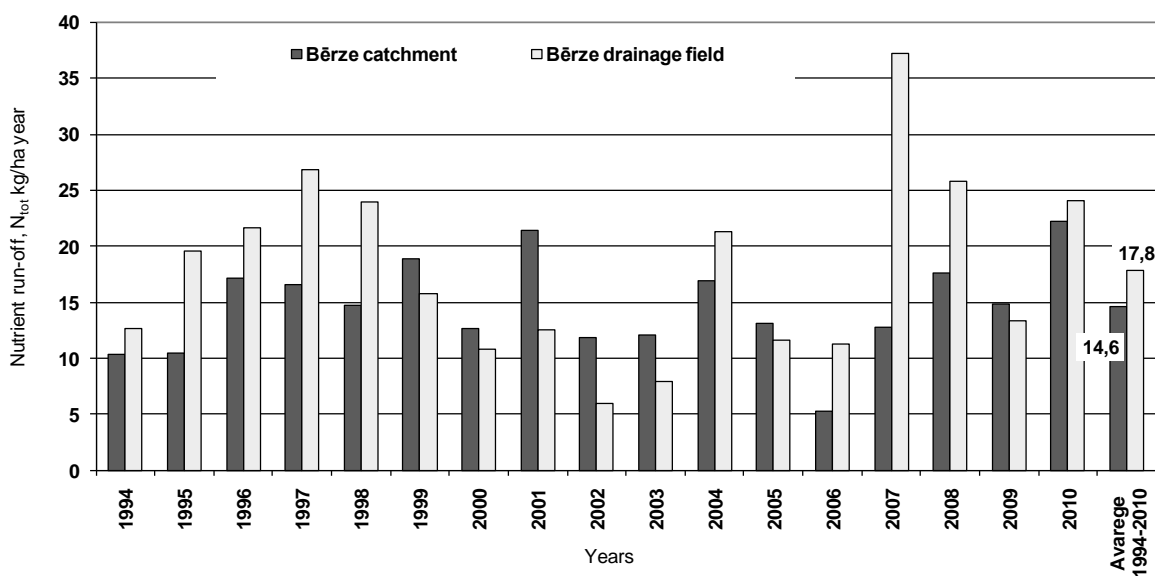


Fig. 9. Nitrogen run-off from diffuse pollution sources, Bērze monitoring station.

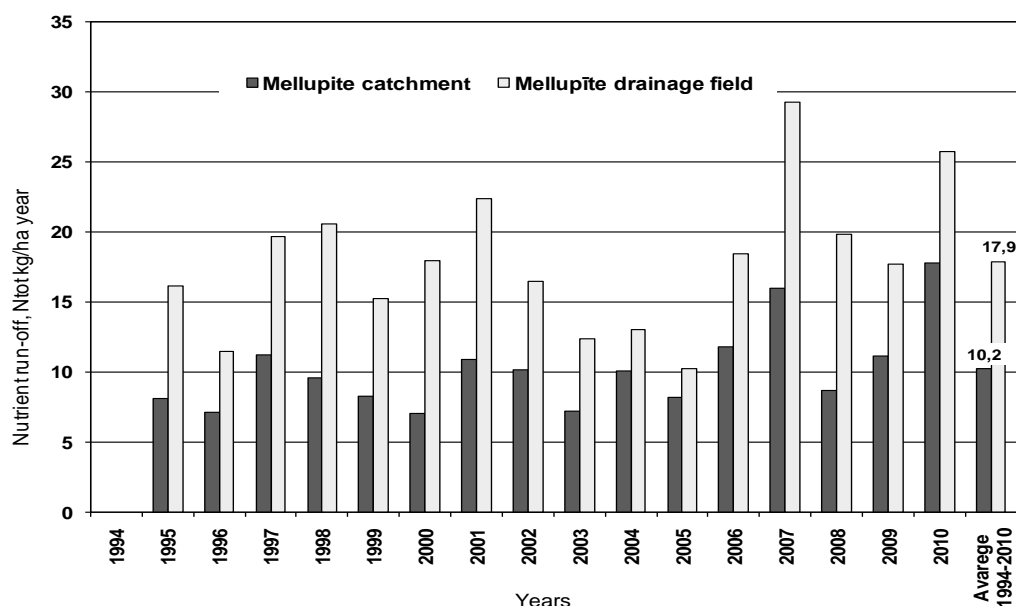


Fig 10. Nitrogen run-off from diffuse pollution sources, Mellupīte monitoring station.

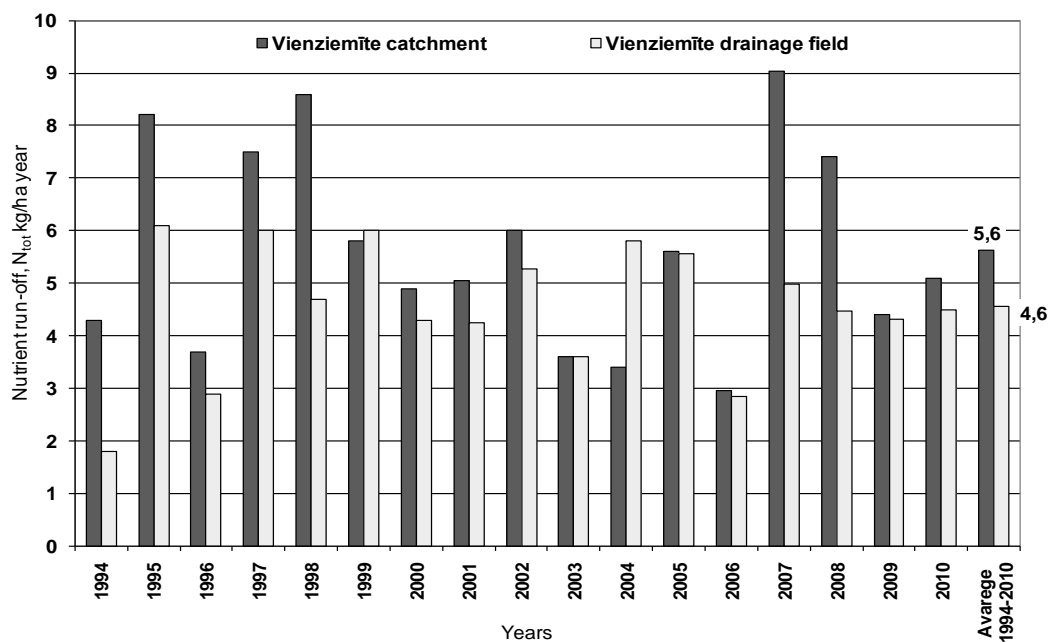


Fig 11. Nitrogen run-off from diffuse pollution sources, Vienziemīte monitoring station.

Our study showed that, compared with diffuse nitrogen pollution, phosphorus run-off has no direct correlation with the intensity of agriculture. In the case of phosphorus, nutrient problems are related to excess precipitation and/or snowmelt which can lead to surface runoff induced erosion and nutrient

loss, causing deteriorating water quality in the field drainage and catchment scale. During 1994 to 2010 the highest phosphorus run-off was observed in the Bērze site (Fig. 12 and 13) during the spring flood.



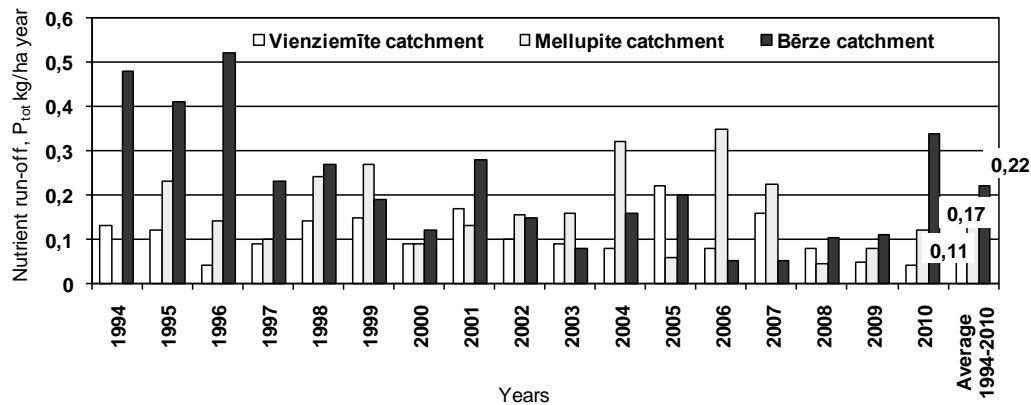


Fig. 12. Phosphorus run-off from diffuse pollution sources in small catchment scale, Bērze, Mellupīte, Vienziemīte monitoring stations.

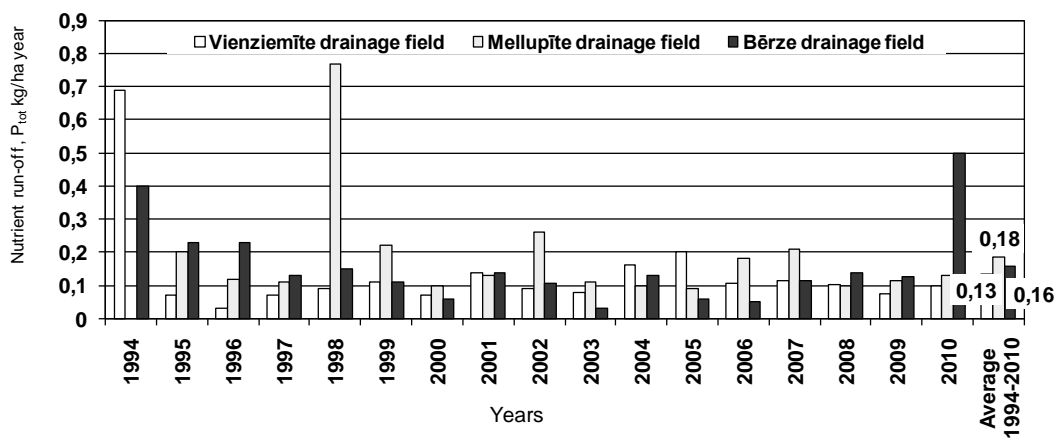


Fig. 13. Phosphorus run-off from diffuse pollution sources in field drainage scale, Bērze, Mellupīte, Vienziemīte monitoring stations.

#### D. Nutrient run-off in the point source pollution monitoring points

Nutrient run-off from agricultural point sources (animal farm) for the Bauska monitoring point is presented in Fig. 14 and 15. Overloading fields with manure from too many animals or using improper manure handling practices caused an average nitrogen leaching of about 2 times higher (49 kg

ha<sup>-1</sup> N<sub>tot</sub> year) than leaching from diffuse pollution catchment. Compared with diffuse pollution catchment, phosphorus run-off was more than 10 times higher, reaching extremely high values (16.9 kg ha<sup>-1</sup> P<sub>tot</sub> year) in 2009. High phosphorus leaching was still observed in the Ogre farm that was closed in 1992 [7, 24].

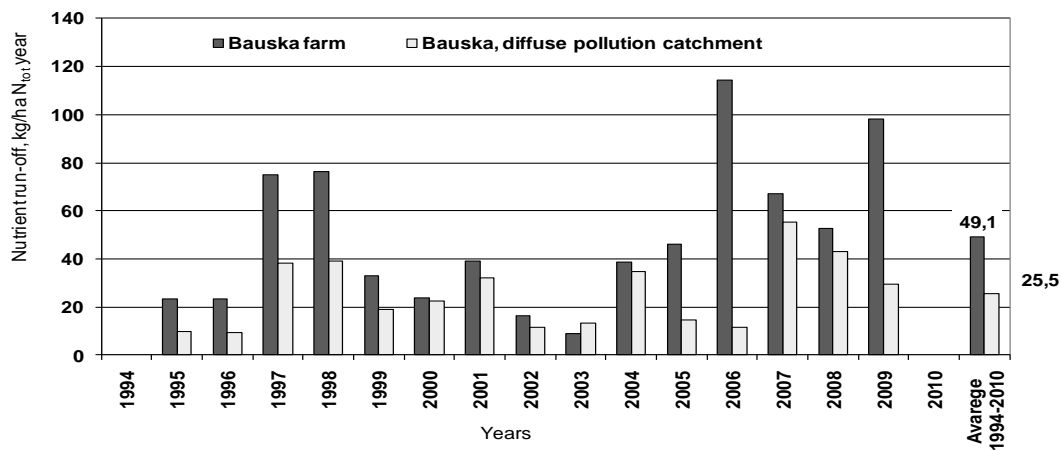


Fig. 14. Nitrogen run-off from point source pollution, Bauska farm monitoring point.

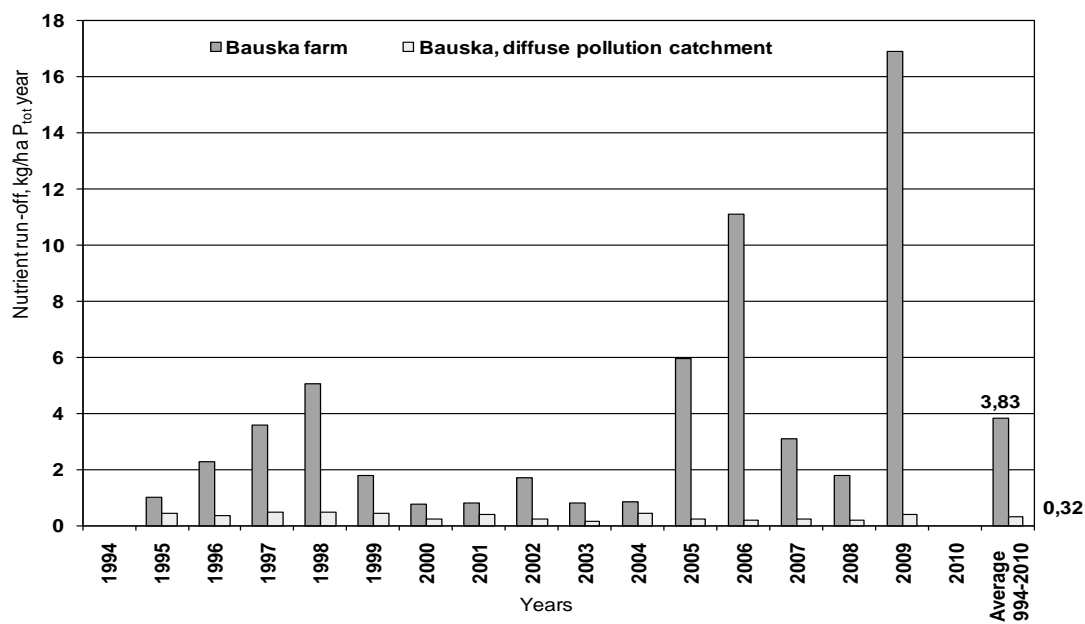


Fig. 15. Phosphorus run-off from point source pollution, Bauska farm monitoring point.

#### IV. CONCLUSIONS

Our research demonstrates how many nutrients may be released into the water environment by leaching in Latvia. It is essential that the relevant monitoring methods and long term monitoring programs are implemented for the assessment of agricultural pollution sources. Due to the climatic conditions, the available data show a high variation in the water quality in both spatial and temporal terms. The total nutrient run-off greatly depends on nutrient concentrations and water discharge, especially in flood periods. On the field drainage scale, nitrate concentrations are often higher than the nitrate limit value set by the Nitrates Directive.

The increased intensity of farming has resulted in an upward trend in both nutrient concentrations and nutrient run-off. Weather-driven hydrology, agricultural land use (acreage of arable land) and fertilizer application are the main factors which impact diffuse source nutrient run-off.

Poor manure handling practices could substantially increase the point source nutrient runoff from agriculture, e.g., as regards to phosphorus leaching, it could be increased more than 10 times, compared with diffuse pollution sources.

National agricultural policy and support measures should integrate agro-environmental requirements to mitigate nutrient pollution of inland water ecosystems.

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**Viesturs Jansons, Ainis Lagzdīns, Laima Bērziņa, Rīvars Sudars, Kaspars Abramenko. Augu barības vielu izskalošanās no lauksaimniecības zemēm Latvijā mainība laikā un telpā: Noplūdes un aiztures ilgtermiņa tendenci mazos sateces baseinos un drenu sistēmās**

Lauksaimniecība var izsaukt paaugstinātu ūdeņu vides piesārņojumu. LLU uzsāka lauksaimniecības piesārņojuma monitoringu 1994.g., izveidojot monitoringa sistēmu lauksaimniecības difūzā un punktveida piesārņojuma noteikšanai. Augu barības elementi – slāpekļa un fosfora savienojumi ūdeņos var nonākt no dažāda rakstura piesārņotājiem. Difūzais piesārņojums dod galveno slāpekļa piesārņojuma daļu Latvijas virszemes ūdeņos. Difūzais piesārņojums ar slāpekļa un fosfora savienojumiem eksistē vienmēr un visur, to var tikai daļēji samazināt ar agro-vides pasākumiem. Punktveida piesārņojums lauksaimniecībā saistīts ar nesakārtotu kūsmēsļu saimniecību lielajās lopkopības fermās. Parasti lauksaimniecības intensifikācija un šaura specializācija noved pie paaugstinātas vides piesārņošanas, kuras novēršanai vajadzīgie finanšu resursi var būt lielāki, kā īslaicīgais ieguvums no ražošanas intensifikācijas. ES un Latvijas likumdošanas akti prasa kontrolēt lauksaimniecības ietekmi uz ūdens vidi, lai samazinātu piesārņojumu ar slāpekli un citiem biogēniem elementiem. Darbā apskatīti ilgtermiņa (1994.g. - 2010.g.) lauksaimniecības piesārņojuma monitoringa rezultāti, kas parāda, ka nitrātu slāpekļa koncentrācijas ir atkarīgas no monitoringa sistēmas līmeņa (drenu lauks, mazais sateces baseins) un lauksaimnieciskās ražošanas intensitātes. Izmantojot ilgtermiņa datu rindas, novērtēta biogēno elementu koncentrāciju mainība un atšķirības drenētos laukos un mazos sateces baseinos. Mazo sateces baseinu līmenī nitrātu robežvērtība visbiežāk tiek pārsniegta Bērzes monitoringa stacijā, plāfībās ar intensīvu lauksaimniecību. Ilgtermiņā novēroto (1994. - 2010.g.) slāpekļa savienojumu tendenci uzrāda koncentrāciju pieaugumu Bērzes un Mellupītes stacijās un nenozīmīgas izmaiņas Vienziemītes stacijā. Bērzes monitoringa stacijā, intensīvas lauksaimnieciskās ražošanas apstākļos, atsevišķos gadījumos (ziemas periods) nitrātu koncentrācijas ievērojami pārsniedza Nitrātu direktīvas noteiktos robežlielumus. Vienziemītes stacijā (ekstensīva lauksaimniecība) slāpekļa koncentrācijas ir zemas un tuvas dabiski tīriem ūdeņiem. Augstākās slāpekļa un, it sevišķi, ūdeņu ekosistēmām bīstamā fosfora koncentrācijas novērotas punktveida piesārņojuma monitoringa postējos Bauskā un Ogrē.

**Виестуре Янсонс, Айнис Лагздиньш, Лайма Берзиня, Ритварс Сударс, Каспарс Абраменко. Изменения во времени и территории вымывания питательных веществ растений из сельскохозяйственных в Латвии: Многолетние тренды выноса и задержки в бассейнах малого водозабора и в дренажных системах**

Сельское хозяйство может способствовать повышенному загрязнению водной среды. ЛСХУ начала мониторинг сельскохозяйственного стока в 1994 году, создавая в Латвии систему мониторинга для определения диффузного и точечного загрязнения от сельского хозяйства. Питательные вещества растений – соединения азота и фосфора поступают в водные объекты от загрязнителей разного характера. Диффузное загрязнение дает основную часть загрязнения азотными соединениями, поступающими в поверхностные воды. Диффузное загрязнение азотом и фосфором существует всегда и везде, его можно только уменьшить мероприятиями по защите агросреды. Точечное загрязнение от сельского хозяйства связано с неполадками в применении органического удобрения в больших животноводческих фермах. Как правило, интенсификация сельского хозяйства и узкая специализация приводит к повышенному загрязнению среды, для устранения которого понадобятся большие финансовые ресурсы, чем временная польза от интенсификации производства. Нормы законодательства ЕС и Латвии для снижения загрязнения азотом и другими биогенными элементами, требуют установить контроль влияния сельского хозяйства на водную среду. В работе приводятся результаты многолетнего (1994 - 2010 гг.) мониторинга загрязнения от сельского хозяйства, которые показывают, что концентрации нитратного азота зависят от уровня мониторинга (бассейн малого водозабора и дренажная система) и уровня интенсификации производства. Используя данные многолетних наблюдений, оценены изменения концентрации биогенных элементов и их различия в бассейнах малого водозабора и в дренажных системах. В бассейнах малого водозабора предельные концентрации нитратного азота часто превышаются на станции мониторинга «Bērze», в угодиях с интенсивным сельским хозяйством. Анализ трендов многолетних наблюдений (1994 - 2010 гг.) показывают прирост концентрации на станциях мониторинга «Bērze» и «Mellupīte» и незначительные изменения в «Vienziemīte». На станции мониторинга «Bērze», в условиях интенсивного сельскохозяйственного производства, в отдельных случаях (зимний период) концентрации нитратного азота существенно превышали предельные концентрации, установленные Нитратной Директивой. На станции «Vienziemīte» (экстенсивное сельское хозяйство) концентрации нитратного азота низкие и близки к фоновой концентрации природных вод. Самые высокие концентрации азота и особенно опасного для водных экосистем соединений фосфора наблюдались в постах мониторинга точечного загрязнения Bauska и Ogrē.